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Soil erosion and non-point source pollution impacts assessment with the aid of multi-temporal remote sensing images

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Abstract

Soil erosion associated with non-point source pollution is viewed as a process of land degradation in many terrestrial environments. Careful monitoring and assessment of land use variations with different temporal and spatial scales would reveal a fluctuating interface, punctuated by changes in rainfall and runoff, movement of people, perturbation from environmental disasters, and shifts in agricultural activities and cropping patterns. The use of multi-temporal remote sensing images in support of environmental modeling analysis in a geographic information system (GIS) environment leading to identification of a variety of long-term interactions between land, resources, and the built environment has been a highly promising approach in recent years. This paper started with a series of supervised land use classifications, using SPOT satellite imagery as a means, in the Kao-Ping River Basin, South Taiwan. Then, it was designed to differentiate the variations of eight land use patterns in the past decade, including orchard, farmland, sugarcane field, forest, grassland, barren, community, and water body. Final accuracy was confirmed based on interpretation of available aerial photographs and global positioning system (GPS) measurements. Finally, a numerical simulation model (General Watershed Loading Function, GWLF) was used to relate soil erosion to non-point source pollution impacts in the coupled land and river water systems. Research findings indicate that while the decadal increase in orchards poses a significant threat to water quality, the continual decrease in forested land exhibits a potential impact on water quality management. Non-point source pollution, contributing to part of the downstream water quality deterioration of the Kao-Ping River system in the last decade, has resulted in an irreversible impact on land integrity from a long-term perspective. © 2005 Elsevier Ltd. All rights reserved.

Keywords: Land use classification; Remote sensing; Environmental impact assessment; River basin management; Erosion control; Non-point source pollution

1. Introduction

With rapid socio-economic changes and various environmental perturbations during the last decade, land resources of populated areas, such as the Kao-Ping River Basin in South Taiwan, have been depleted and eroded significantly, resulting in increased ecological vulnerability and hydrological disruption. Furthermore, the instable geological structure resulting from earthquake impacts has led to accelerated soil erosion, increased turbidity of rivers, and

more incidences of debris flow when storm events occur. Ecosystem integrity due to such variations also becomes an urgent focus for the prospective planning of the land exploitation in this river basin. Change detection of land use and land cover is one of the essential practices in many interrelated disciplinary areas, such as soil erosion, flood control, landscape conservation, ecosystem restoration, and water quality management via non-point source pollution control. Careful monitoring and assessment with different temporal and spatial scales in this regard would reveal a fluctuating interface, punctuated by changes in rainfall and runoff, movement of people, perturbation from environ-mental disasters, and shifts in agricultural activities and cropping patterns.

Many previous analyses focused on investigating the 111 interactive relationship between soil erosion associated with 112

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non-point source discharge and ecosystem management. 113 Some pinpointed the effects of environmental changes that 114 would impact both the ecological characteristics of plant 115 species and their distribution and abundance in specific 116 landscapes (Hoffmann, 1998). Others assessed landscape 117 changes in terms of landscape functions and conservation 118 119 potential to form a comparative basis in the search for the optimal ecological management alternative (Bastian and 120 Roder, 1998). The most susceptible factor that affects 121 ecosystem functioning is actually the seasonal patterns of 122 soil moisture distribution, inorganic nutrients, organic 123 nitrogen, and soil erosion over the most abundant vegetation 124 types in some typical watersheds around the world 125 (Arhonditsis et al., 2000). In the past, environmental studies 126 were often designed to determine the nutrient concentrations 127 in runoff sediments in order to quantify the outflows of the 128 ecosystem's nutrient budget for preserving agricultural 129 productivity and diminishing non-point pollution (Arhon-130 ditsis et al., 2000). Some case studies explored socio-131 economic, climatic, and lithological components of the 132 erosion processes leading to development of a risk-based 133 134 map associated with the large-scale clearance of dispersive soils for arboriculture (Warren et al., 2001; Faulkner et al., 135 2003). 136

Best management practices (BMPs) have long been 137 recognized as an integral part of water pollution prevention 138 and control in river basins for controlling non-point source 139 pollution using ecologically benign approaches. Simulation 140 analyses of non-point source pollution impacts, which are 141 instrumental to Total Maximum Daily Loads (TMDLs) 142 programs, play an important role in decision-making 143 (Shoemaker et al., 1997). They help planners identify, 144 analyze, and simulate the impacts of alternative land use 145 management policies and practices with respect to non-point 146 source pollution control. Various types of non-point source 147 numerical simulation models were employed to account for 148 149 the integrated impacts of hydrological cycle and land use pattern in relation to nutrient yield (Bailey et al., 1974; 150 Donigian et al., 1996). They can be further classified as 151 distributed (i.e. grid-based approach) and lumped (i.e. semi-152 mechanistic approach) parameter hydrologic models. Two 153 salient examples based on a monthly scale in each category 154 include the generalized watershed loading function (GWLF) 155 (Haith et al., 1992) and the Cornell non-point sources 156 157 simulation model (CNPS) (Dikshit and Loucks, 1995, 1996). More elaborate grid-based numerical simulation models for 158 assessing non-point source waste loads in the agricultural 159 field may enable planners to assess detailed pros and cons of 160 different policy options. They generally simulate rainfall, 161 soil erosion, run-off sediment, temperature, wind speed, 162 atmospheric pressure, and non-point source processes 163 leading to estimate pollutant loading at the watershed outlet. 164 165 Existing examples include ANSWERS (Beasley and Huggins, 1982), HSPF (Donigian et al., 1980), and AGNPS 166 (Young et al., 1989). Synergy of different models is 167 sometimes anticipated for designated applications with 168

various temporal and spatial scales. For instance, the 169 Spatially Integrated Models for Phosphorus Loading and 170 Erosion (SIMPLE) model was developed to evaluate the 171 potential phosphorus loading to streams from areas with 172 various soil and management practices. This model operates 173 on a daily time step and independent simulations are based on 174 factors such as rainfall, soil characteristics, fertilizers and 175 animal waste applications, and topographic characteristics 176 (Kornecki et al., 1999). It is often connected with the Erosion 177 Productivity Impact Calculator (EPIC) which is a physically 178 based model designed to simulate the effect of different 179 management practices on crop yield and on chemicals, 180 including phosphorus losses by surface runoff, sediment 181 movement, and leaching below the root zone to accomplish 182 an integrated study. Recent efforts using a neural network 183 model demonstrate an additional dimension for soil erosion 184 modeling (Licznar and Nearing, 2003). With the aid of 185 various environmental models, the improvement of the 186 estimation and control of non-point sources in some reservoir 187 watersheds was greatly enhanced in recent years (Safe and 188 Choudhury, 1998; Yool, 1998; and Miller et al., 1998). 189

The integration of grid cell information in the study area 190 of interest in geographic information systems (GIS) with 191 various environmental models has been fully implemented 192 (Liang and Chen, 1995; Goodchild et al., 1996; Dikshit and 193 Loucks, 1995, 1996). In many cases the grid-based non-194 point sources modeling analysis is prohibitively hindered 195 because of insufficient information and unbearable data-196 intensive requirements. However, land use characterization 197 and change detection analysis based on remote sensing are 198 able to provide planners with sufficient background 199 information for model parameterization (Helmschrot and 200 Flügel, 2002). Recent applications with different spatial 201 scales range from a continental scale to a river basin scale 202 and to an urban scale (Cohen et al., 2003; Hashiba et al., 203 2000; Tapiador and Casanova, 2003). With such a need, 204 advanced and improved remote sensing data analysis and 205 land use classification techniques are anticipated (Steele, 206 2000; Foody, 2002; Pal and Mather, 2003). 207

The goal of this study is to assess the long-term impact of 208 soil erosion and nonpoint source pollution in a fast growing 209 river basin in Taiwan via the integrative use of remote 210 sensing, global positioning system (GPS), GIS, and nume-211 rical simulation models. While a companion study (Ning et 212 al., 2002) addressed available data, choices of method, and 213 what has been done with these data in a short-term analysis 214 for the Kao-Ping River Basin, this paper focuses on the 215 application of multi-temporal remote sensing images to aid 216 in soil erosion and non-point source pollution assessment 217 within the same study area from a long-term perspective. 218

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2. Study area

The Kao-Ping River Basin is located from 120° and $22 \text{ min } (122^{\circ}22')$ north longitude and 22° and 28 min

 $(22^{\circ}28')$ east latitude to 121° and $3 \min (121^{\circ}3')$ north 225 longitude and 22° and $30 \min (22^{\circ}30')$ east latitude. The 226 mainstream flows through approximately 140 km and drains 227 228 towards Southern Taiwan Strait. With an area of 3256 km², 229 the mainstream of the Kao-Ping River originates from four 230 small tributaries: Chi-San River, Liao-Nung River, Cho-231 Kou River, and Ai-Liao River (Fig. 1). From the confluence 232 to the union with those tributaries at the location of Li-Ling 233 Bridge, the river carries the name Kao-Ping River. The river 234 basin elevations range from sea level near the estuary region 235 to 3293 m upon the headwater with the average slope of 236 about 1/150. Five soil types appear in this river basin.

Entisols, which are Lithosols, are spread around the 281 northern and eastern mountainous area. Inceptisols and 282 Alfisols that come from the older alluvial soils are 283 distributed over the southwestern plain. Ultisols and 284 Oxisols, which constitute the red soils, cover the southern 285 region. Inceptisols and Ultisols that represent most yellow 286 soils are dominant soils around the midwestern area. The 287 288 land use pattern is quite distinct. Forested land covers the 289 northeastern mountainous region in the upper portion of the 290 river basin. Residential districts are mostly located down-291 stream in this river basin. While many groves are located 292 along the river corridor, agricultural land spreads around



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the western portion of the river basin. There are also a
number of small- and medium-scale industrial parks in the
populated region downstream.

340 The mean annual rainfall in this river basin is close to 3000 mm, over 90% of which precipitates in the wet season 341 342 from May to September. The period of high flow rate in the stream usually occurs in late spring and summer due to the 343 impacts of monsoons and typhoons. In this time period, the 344 Kao-Ping River flow increases to a level approximately 8-345 346 12 times higher than the dry season flow. Uneven rainfall 347 over seasons has resulted in severe issues of water resources 348 redistribution in the winter and early spring that inevitably requires building more reservoirs for water storage. Land 349 350 subsidence occurred in some coastal areas due to the abuse 351 of groundwater resources.

352 The drainage area of the Kao-Ping River Basin is 353 primarily planned for agricultural production. Crops that are 354 produced from the agricultural fields include rice, sugar-355 cane, pineapple, and a variety of vegetables. Stock farming 356 is an active agricultural activity. In addition to the use of 357 water for agricultural production and industrial manufactur-358 ing processes, water is also essential for drinking and 359 personal hygiene in this area. The Kao-Ping River system, 360 however, has a long history of higher BOD and NH₃-N 361 concentrations due to careless landfill operations, 362 inadequate disposal of manure from stock farming, and 363 continuous discharges of industrial and domestic waste-364 water effluents. These discharges in the middle and 365 downstream areas of the Kao-Ping River system, where

most water intakes are located, have resulted in acute needs 393 for promoting a new management policy to improve the 394 water quality condition. The attainable use of river water is 395 officially classified into three categories for management 396 and control purposes. Several pollution prevention pro-397 grams have been put into practice by the government agency 398 (Ning et al., 2001). These include a large-scale livestock 399 subsidy program for removing almost half of the domestic 400 livestock in the upstream and middle stream areas and a 401 series of collective sewer construction projects in several 402 cities along the river corridor. On the other hand, controlling 403 the agricultural run-off would require a complete assess-404 ment of soil erosion and associated non-point source 405 pollution impacts in the river basin from a long-term 406 perspective. 407

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3. Methods

412 The analysis of soil erosion associated with non-point 413 source pollution impacts includes three stages (Fig. 2): (1) 414 land use pattern classification, (2) data acquisition for 415 modeling analysis, and (3) simulation analysis for non-416 point source pollution impact assessment. The classification 417 of land use pattern is viewed as one of the prerequisites for 418 applying the subsequent simulation analysis. Thus, the initial 419 focus of this study was on classifying eight types of land-use 420 patterns in the watershed with the aid of SPOT satellite 421 images that were generated during the time period of



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1991-2001. A change detection analysis of decadal land use 449 patterns in the river basin was then performed. To support the 450 impact analysis of soil erosion and associated non-point 451 sources, it requires a series of essential databases integrating 452 both remote sensing and non-remotely sensed data for 453 modeling analysis. In the end, the GWLF model provides 454 predictions of monthly nitrogen and phosphorus loads in 455 stream flow within the time frame based on the given rainfall 456 457 and runoff patterns and the projections of soil erosion and 458 sediment transport. They are described in detail below. 459

460 461 3.1. Land use pattern classification

462 Many shifting land use patterns, driven by a variety of 463 social causes, result in land cover changes that affect 464 biodiversity, water balance, radiation budgets, trace gas 465 emissions, and other processes that, cumulatively, affect the 466 global climate and biosphere (Riebsame et al., 1994). A 467 primary component of mapping land use and land cover is to 468 develop a land use classification system. The supervised 469 classification approach was applied in the land use pattern 470 classification of this study. It needs to be carried out in 471 sequence including satellite image mosaic, classification, 472 and verification. 473

474 Supervised image classification is a method in which the 475 analyst defines training sites on the image that are representative of each desired land cover category. The 505 delineation of training areas representative of a cover type is 506 most effective when an image analyst has knowledge of the 507 geography of a region and experience with the spectral 508 properties of the cover classes (Skidmore, 1989). The image 509 analyst then trains the software to recognize spectral values 510 or signatures associated with the training sites. After the 511 signatures for each land cover category have been defined, 512 the software then uses those signatures to classify the 513 remaining pixels (ERDAS, 1999). Fig. 3 indicates the 514 detailed steps of this procedure. They can be described using 515 516 a SPOT satellite image as follows:

3.1.1. (1) Select and mosaic satellite image

519 At first, an overlay is used when it is desirable for a block 520 of content of a SPOT image and GIS themes of the study 521 area to be shared with respect to the 2°-Zone Transverse 522 Mercator (TM) projection coordinates. The SPOT has high-523 resolution visible (HRV) imagery, which includes three 524 bands with 20-m resolution color mode and 10-m resolution 525 panchromatic mode. Colorful images include infrared, red, 526 and green bands corresponding to the wavelengths of 527 $0.79 \sim 0.89 \,\mu\text{m}, 0.61 \sim 0.68 \,\mu\text{m}, \text{ and } 0.50 \sim 0.59 \,\mu\text{m},$ 528 respectively. Satellite image reception, archiving, and 529 validation were carried out by the Center for Space and 530 Remote Sensing Research in Taiwan. Current HRV data 531



precision (80 bit) appears suitable for land cover identifi-cation on a river basin scale.

The entire analysis for land use identification and 563 classification in the watershed of the Kao-Ping River 564 Basin was designed based on a practical scale GIS 565 566 framework. The study area has been further classified by 567 the authors using three bands of SPOT satellite images in the 568 spectrum that were collected during the time periods of 569 1994 (i.e. November, 1994), 1996 (i.e. June, 1996 and 570 August, 1996), 1997-1998 (i.e. November, 1997 and 571 January, 1998), and 2001 (i.e. January, 2001). Fig. 4 572 shows the decadal SPOT satellite images of the Kao-Ping 573 River Basin. These SPOT scenes selected for land use 574 pattern identification covering the portions of the Kao-Ping 575 River Basin in South Taiwan must be first verified to a status 576 of GICS level 10 by the Center for Space and Remote 577 Sensing Research before performing various environmental 578 applications. Automatically matching conjugate features in 579 overlapped images was required first. Due to time 580 differences when producing these satellite images, direct 581

matching of them was impractical. Instead of using advanced direct matching algorithms, this study applied a different method in which each satellite image acquired at a different time was used for land use classification independently and individual outcomes were then integrated together as a whole for final evaluation via the use of Erdas Imagine[®] image processing software.

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3.1.2. (2) Classification of land use patterns

Application of GPS helps verify the effectiveness of land 627 use classification based on SPOT satellite image. The 628 information gained from GPS can be divided into two 629 groups. One group is prepared as a set of feature points that 630 is designed for a direct calibration of land use classification 631 in the Erdas Imagine[®] image processing system. The other 632 group is prepared as an additional set of feature points or 633 ground-control points (GCPs) for validation purposes. Fig. 5 634 shows the locations of all GPS feature points prepared for 635 ground truth verification in 2001. Partial feature points 636 expressed as dots were used for supervised land use 637



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Fig. 5. The distribution of feature points for calibration (2001).

classification, while those expressed as triangles were saved for final validation of land use patterns. The overall supervised ground truth classification process can therefore be trained based on the first set of feature points in the first stage and then be validated by the given land use pattern in the watershed with the aid of the second set of feature points in the second stage. An independent survey of land utilization in 1991, as shown in Fig. 6, was selected as the baseline information to track down the variations within a ten-year time frame when looking for the temporal variations of land use on a long term basis.

3.1.3. (3) Verification of land use pattern by local aerial photographs

Final accuracy can be confirmed by using some available aerial photographs as a kind of supplementary ground truth verification. A geometrical rectification procedure is required to map the aerial photographs onto the ground coordinate system. In this paper, the aerial photographs were rectified using an ortho-rectification procedure. The original hardcopy photographs were digitized first for digital processing. The photogrammetric procedure of Digital Aerial Triangulation was then performed to recover the exterior orientation. Because the correction for terrain displacement requires a Digital Surface Model (DSM), automatic stereo image matching was carried out to generate DSM. Then, the images were ortho-rectified



Fig. 6. The surveyed land use patterns in the Kao-Ping River Basin (1991)

according to the perspective geometry and terrain displacement. Finally, all rectified images were mosaiced together to compose an overall rectified image, which enabled us to overlap the aerial image on the satellite images. This eventually led to the accuracy assessment for the outputs of land use pattern classification. To achieve this goal, Kappa statistics are normally selected to express the proportionate reduction in error generated by a classification process, compared with the error of a completely random classification (Congalton, 1991). Fig. 7 indicates the area picked up for accuracy assessment points, in which a stratified random sampling method was applied. It uniquely allows us to evaluate a classified image file for evaluating diagnostic imaging methods.

3.1.4. (4) Change detection analysis of decadal land use in the designated river basin

An increasingly interesting application of remotely sensed data in the context of land use pattern classification is for change detection. Change detection is an important process in monitoring and managing land resources because it provides quantitative analysis of the spatial variations of the land use patterns within the designated time frame. Based on the remote sensing efforts aforementioned, this paper is also designed to summarize all the spatial variations

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828 Table 1

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The nutrient concentration applied in this analysis (Ning et al., 2002)

Land use patterns	Forest	Community	Orchard Ag	gricultural land	Grassland	Sugarcane
TN (mg/L)	1.925	1.806	1.709 17	7.74	0.3	10.69
TP (mg/L)	0.18	1.17	0.46 9	0.56	0.15	4.66
Nutrient content in t	he sediment					
Soil classification	Podzolic soils Red soils	Mudstone, silt-	Mudstone, silt-shale stone			Regosol
		Non-weatherin	g Semi-weather	ring Weat	hering	-
			126	207		318
TN (mg/kg)	877	78	136	597		010

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terrain modeling (DTM) information, mode of cover and management types for handling field crops, and the condition of agricultural practices in the field. But the representation of flow directions and calculation of upslope areas using rectangular grid digital elevation models must be applied (Tarboton, 1997). Finally, the mean concen-trations (i.e. export coefficients) of nutrients in runoff and sediment transport from both rural and urban regions were collected and combined based on several previous sampling practices in Taiwan (Ning et al., 2002). They are

summarized as part of the semi-empirical inputs in this 953 model (see Table 1) (Ning et al., 2002). 954

3.3. Simulation analysis

The GWLF model was selected as a baseline tool to estimate the non-point source waste loads reflecting the impacts from soil erosion processes and nutrient cycles in the watershed of the Kao-Ping River Basin. The GWLF model describes non-point sources for runoff, erosion and



1009 urban wash off, and a lumped parameter linear reservoir groundwater model. The delivery ratio concept was 1010 defined based on the Runoff Curve Numbers and the 1011 Universal Soil Loss Equation methods (Vanoni, 1975). It 1012 helps to yield an estimation of non-point source waste 1013 loads with respect to soil erosion, sediment transport, and 1014 resulting nutrient fluxes, such as total phosphorus (TP) 1015 and total nitrogen (TN). Eight types of land use patterns 1016 that have been classified in the previous step can be 1017 incorporated to track down long-term non-point source 1018 impacts. 1019

All data from the SPOT satellite images were eventually 1020 integrated into the ArcView® system, in which land use and 1021 land cover information can be extracted from the GIS 1022 database to help in the estimation of soil loss and non-point 1023 1024 pollutant loading in the watershed. The GIS expresses the geographic area as a matrix of grid cells of which each grid 1025 cell stands for a parcel of land of certain size and serves as the 1026 basic unit for quantification of various physical, topographi-1027 cal, meteorological, hydrological and geological features. 1028 Thereafter, the spatial analyst module in GIS can be applied 1029 1030 for the prediction analysis with respect to each type of land use pattern in the watershed in a way that enables us to 1031 estimate individual impacts due to each type of non-point 1032 source of interest. Total impacts of soil erosion and non-point 1033 sources can then be integrated with respect to all land use 1034 patterns identified by SPOT satellite images and associated 1035 GIS operations. 1036

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1039 4. Results and discussion 1040

1041 4.1. Assessment of long-term land use variations 1042

1043 The spatial analyst module embedded in the ArcView[®] 1044 system enables us to compute the land use statistics 1045 corresponding to each drainage sub-basin. Except for the 1046 area covered by clouds in the satellite image, the land use 1047 patterns being categorized include orchard, farmland, 1048 sugarcane farm, forest, grassland, barren, community, and 1049 water body in each sub-basin. These land-use patterns can 1050 be delineated as an ArcView[®] cartographic output as shown 1051 in Fig. 8. 1052

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Class name	Reference points	Kappa coefficient
Orchard	78	0.8449
Forest	27	0.6917
Grassland	12	0.8670
Community	5	0.2308
Farmland	29	0.8051
Sugarcane	2	1.0000
Barren	44	0.8397
Water body	3	0.7462
Total	200	0.7648

For the sake of further assessment of classification 1065 accuracy, GPS feature points and aerial photographs were 1066 used as real ground truth for final confirmation. Table 2 1067 shows the number of points in the assessment and the 1068 associated Kappa coefficient for each land use pattern. The 1069 Kappa agreement test was applied and an average accuracy 1070 of 0.7648 was achieved with respect to different Kappa 1071 coefficients associated with different land use patterns. 1072

Table 3 summarizes land use patterns on a comparative 1073 basis and covers the time period from 1991 to 2001. The 1074 survey of land use patterns in the year 1991 was manually 1075 digitized for comparison. At present, forested land occupies 1076 approximately 60.3% of the total area of the river basin. It 1077 1078 spreads out along the eastern mountainous area in the watershed. Orchard areas that keep increasing over time are 1079 1080 scattered around on both sides of the river corridor. They 1081 occupy approximately 17.7% of the total area in this river 1082 basin. Farmland, grassland, barren, and community constitute approximately 9.9, 3.3, 2.2 and 4.1% of the total area, 1083 1084 respectively. They are normally located at the southwest 1085 region of the river basin. Because barren was not included in 1086 the classification, the dry riverbed was classified as part of 1087 the water body, which results in some inaccuracy. The 1088 stream flow rate in the wet season may be ten times higher 1089 than that in the dry season. Such seasonal stream flow 1090 variations were so phenomenal that they increased 1091 classification inaccuracy with respect to the water body. 1092 Due to the effect of clouds, inevitable discrepancies arise 1093 from the adoption of different satellite images acquired in 1094 the wet seasons in 1994 and 1996 and the dry seasons in 1095 1998 and 2001. On the other hand, the increasing with-1096 drawal of water flows from the river to satisfy the rising 1097 demand for water resources over time could affect the area 1098 of the water body (Ning et al., 2002). In 1998, clouds 1099 covered an area of about 62 km², which also impacts the 1100 classification of forested land in the watershed. 1101

Fig. 9 summarizes the decadal variations of land use patterns in the Kao-Ping River Basin. From 1994 to 2001, the land use for orchard had increased dramatically due to

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Land-use patterns	1991	1994	1996	1998	2001
Orchard	132.44	385.90	389.60	471.44	576.22
Forest	2242.30	2182.28	2048.55	1816.97	1962.52
Grassland	169.48	114.53	132.68	206.51	106.99
Commu- nity	69.74	71.96	108.55	153.44	133.59
Farmland	442.84	267.80	285.71	255.17	323.10
Sugar- cane	102.06	47.26	47.91	46.19	45.89
Barren	0.00	132.98	191.60	207.06	70.70
Water body	91.71	48.72	41.97	38.08	29.30
Cloud	0.00	0.86	9.80	61.71	3.99
Total	3250.58	3252.30	3256.38	3256.57	3252.30

Table 3

¹⁰

fast economic growth region wide. The capacity for 1121 production of versatile fruits contributed to both domestic 1122 and foreign target markets from which the income supports 1123 1124 the local socio-economical system basin wide. In addition, the local pork industry, which counts heavily on stock 1125 farming, has enormous implications for the international 1126 trade surplus due to steady export demand. As a result, the 1127 total area of grassland and community also presented an 1128 increasing trend over time in the last decade. Yet forested 1129 1130 land decreased substantially in the last decade. Fig. 10(a)explains the distribution of forest versus non-forested land 1131 1132 in Taiwan. However, non-forested land only occupies 41.47% of the total area. Fig. 10(b) reveals observational 1133 evidence that while the per capita GDP and the export 1134 volume of the timber remained the same within the last 1135 1136 decade, the reforestation efforts decreased over time in the same time period. A considerable area of forested land 1137 1138 might be eradicated and transferred into golf courses, 1139 recreational parks, orchard farms, tea farms, and residential 1140 areas. The destabilization of fragile mountain slopes, 1141 through deforestation, agricultural expansion, excessive 1142 and indiscriminate grazing and expansion of the road 1143 network, disrupted the hydrological cycle in the river basin. 1144 It also caused the ecosystem irretrievable changes, such as 1145 vegetation area reduction and protozoan and protophyte 1146 disappearance affecting biodiversity in the long run. The 1147 intelligent use, extraction, conservation, and reuse of forest 1148 resources, and the design of long-term strategies require not 1149 only comprehensive understanding of the changes in the 1150 forested land, but also new and appropriate technology and 1151 solutions. To identify the susceptible spots for managerial 1152 use, Fig. 11 further pinpoints the long-term variations in 1153 forested land in the Kao-Ping River Basin (1994-2001). 1154 Overall, the long-term change detection analysis in this 1155 study showed that the land development programs in the last 1156 decade significantly changed the land use patterns in this 1157 region such that the land management, ecological conserva-1158 tion, and pollution control actions have become a new 1159 challenge to the environmental decision makers in the river 1160 basin.







Fig. 10. The statistics of forest change versus timber export in Taiwan (1993–2002).



Fig. 11. The variation of forest regions in the Kao-Ping River Basin (1994–1231 2001). 1232

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4.2. Assessment of soil erosion and non-point sourcepollution impact

1236 Simulation outputs using the GWLF model reveal that 1237 more than ninety percent of the total impacts of TN and TP appear from June to September within a year when rainfall 1238 1239 goes up in the summer. The long-term change detection 1240 analysis with regards to soil erosion and associated non-1241 point source pollutant loading between years can also be 1242 explored based on the land use patterns in different years. 1243 Table 4 summarizes the decadal changes in total nutrient 1244 fluxes and soil erosion impacts. The impacts of non-point 1245 source pollution over the years result from a complex 1246 environmental process due to movement of people, changes 1247 in land use patterns, and shifts in agricultural activities and 1248 cropping patterns. The positive correlations between rainfall 1249 and soil erosion obtained quantitatively in the analysis may 1250 further verify the complex, dynamic, and interdependent 1251 relationships between them. 1252

The combination of Figs. 12 and 13 would enable us to 1253 graphically demonstrate the possible cross linkages of 1254 spatial and temporal changes in soil erosion associated 1255 with non-point source pollution in the study area. While 1256 Fig. 12 delineates the temporal variations in soil erosion 1257 associated with varying nutrient loading in the last decade, 1258 Fig. 13 further categorizes the distribution of non-point 1259 source pollution in terms of different land use patterns 1260 using the year 2001 as a representative case. It clearly 1261 shows that the levels of annual soil erosion fluctuate 1262 stepwise in association with the variations in annual 1263 precipitation, and are associated with continuous land 1264 cultivation over the years. Within the seven identified land 1265 use patterns, excluding water body, in 2001, farmland is 1266 recognized as the most important one contributing to the 1267 biggest portion of nutrient sources in the river basin 1268 1269 owing to a great deal of manure spreading. This also resulted in a critical situation for water quality manage-1270 ment since most of those farms are situated alongside of 1271 the Kao-Ping River corridor. Furthermore the total area of 1272 1273 forested land has been gradually decreasing, accompanied 1274 by increasing orchard and farmland. As a consequence, 1275 nutrient flux is anticipated to increase as intensive 1276 applications of fertilizers in most agricultural regions 1277

1278 Table 4

1279 The estimate amounts of nutrient loading and soil erosion in the Kao-Ping1280 River Basin

Year	Rainfall (mm)	TN (ton/year)	TP (ton/year)	Erosion (million-ton/ year)
1991	2601.4	1407.6	260.0	213.2
1994	3349.3	2021.0	332.4	275.3
1996	2329.7	1935.0	310.3	264.9
1998	3017.3	2070.8	348.3	289.2
2001	3769.3	2386.1	451.1	356.0



Fig. 12. The variation of nutrient loadings and soil erosion in the Kao-Ping River Basin.

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and inadequate disposal of animal waste in livestock farming communities continue.

4.3. Management policy

1309 Risk maps of nutrient loading and soil erosion impacts 1310 may be produced for essential assessment and policy 1311 decision-making (Khawlie et al., 2002). To ease the 1312 management efforts, Fig. 14 exhibits the ranking of risk of 1313 nutrient loading and soil erosion impacts in the Kao-Ping 1314 River Basin. It involves using three regional patterns, 1315 consisting of high, middle, and low levels, to address how 1316 the resultant changes in land use patterns affect people in 1317 the coupled human and natural systems. While a 1318 significant number of non-point sources are spread around 1319 the middle stream and downstream areas, the regions with 1320 critical soil erosion are located at the upland area in the 1321 watershed. In particular, these regions with high-level 1322 potential for soil erosion are mostly located at the eastern 1323 and northern mountainous areas where the slope is steeper 1324 and the local rainfall is relatively higher than the other 1325 regions. Agricultural activities for crop and orchard 1326 farming have already been extended to those marginal 1327 and sub-marginal areas nearby the forest without taking 1328 the sustainability issue into consideration. This will 1329 eventually lead to the consequences of increased sediment 1330 transport in the river channel and impact the ecosystem 1331



Fig. 13. The non-point sources distribution in the Kao-Ping River Basin in 2001.

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Fig. 14. The potential of nutrient loading and soil erosion in the Kao-Ping
 River Basin.

1361 integrity in the estuary region. To handle the interrelated challenges, the non-point source critical map analysis 1362 1363 obtained from this multidisciplinary study would allow 1364 decision makers to identify, describe, rank, and then 1365 visualize site options in collaborative spatial decision 1366 making. A 'consensus map' is thus devised for future 1367 decision making to visualize both the priority of the 1368 critical area as well as the consensus status about those 1369 priorities in a management sense, which should be 1370 instrumental to land development and environmental 1371 management programs (Sivertun and Prange, 2003).

1372 Parks, including golf courses, designed for recreational 1373 use, could be another vital non-point source in some of 1374 the middle and upstream areas. The continuous develop-1375 ment of golf courses is thus deemed very important in 1376 non-point source pollution control in the context of land 1377 use management. In the presence of so many golf courses 1378 in this region, advanced remote sensing images, such as 1379 IKONO images, should be used to access the correspond-1380 ing non-point source impacts in the future. On the other 1381 hand, it is also worthwhile paying attention to the 1382 pesticides discharged from agricultural practices. The 1383 methods discussed in this study, such as GIS, remote 1384 sensing and the GWLF model, are not suitable for 1385 estimating the impact of pesticides.

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With the rapid increase in population during the last decade in the Kao-Ping River Basin, the demand for land and food has increased considerably, putting increased biotic stress on the critical environmental components, like soil, water and forests. Consequently, this human transformation process has brought about drastic changes in the resource-use practices and land-use pattern of the region.

This study builds on, supports, and integrates many existing
disciplines and tools and provides new insight for river basin
management. It quantifies the decadal changes in land use

patterns via the use of multi-temporal remote sensing 1401 images and non-point source numerical modeling, thereby 1402 providing a firm basis for assessing the impacts of soil 1403 erosion and non-point source pollutant loading in the 1404 coupled human and natural system. It involves consider-1405 ation of the natural and human dimensions of environmental 1406 changes, including different configurations of socio-econ-1407 omic systems, changes in rainfall and runoff, movement of 1408 people, and shifts in agricultural activities and cropping 1409 patterns. 1410

Research findings indicate that while the decadal 1411 increase in orchards poses a significant threat to environ-1412 mental quality, the continual decrease in forested land has a 1413 potential impact on non-point source pollution. Land 1414 degradation along with soil erosion and non-point source 1415 pollution, contributing to part of the downstream water 1416 quality deterioration of the Kao-Ping River system in the 1417 last decade, has resulted in an irreversible impact on the 1418 natural system. It can be concluded that sustained 1419 monitoring with respect to ecosystem features, environ-1420 mental quality indices, and climate change symptoms using 1421 remote sensing should be a meaningful tool to aid in various 1422 land use management programs. This multidisciplinary 1423 approach using remote sensing to aid in decision making 1424 can save time when planners are required to assess the land 1425 use management policy with respect to soil erosion and non-1426 point source pollution impacts. Such information should 1427 prove useful to watershed managers in the nexus of 1428 sustainable development. 1429

j.,	Uncited	reference

Chang and Shaw, 2000.

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